# THE DISTRIBUTION OF VESICULAR-ARBUSCULAR MYCORRHIZAE ON MOUNT ST. HELENS, WASHINGTON

JONATHAN H. TITUS<sup>1</sup>

Department of Botany, Box 355325, University of Washington, Seattle, WA 98195

ROGER DEL MORAL

Department of Botany, Box 355325, University of Washington, Seattle, WA 98195

SHARMIN GAMIET

Department of Botany, Box 355325, University of Washington, Seattle, WA 98195

### ABSTRACT

Vesicular-arbuscular mycorrhizae (VAM) occur in most terrestrial ecosystems and are crucial to understanding community structure and function. However, their role in primary succession is poorly understood. This study examined the distribution of VAM propagules, spores, and plants across the Pumice Plains of Mount St. Helens.

VAM colonized plants and propagules were common in sites with thick vegetation, such as areas of relict pre-eruption vegetation and lupine patches, but were very infrequent in barren areas which comprise nearly all of the Pumice Plain. The vegetation of the Pumice Plain is composed primarily of facultatively mycotrophic species which are currently nonmycorrhizal. Mycorrhizal plants occur in refugia and thickly vegetated areas. VAM spore density and richness was low and spores are essentially restricted to densely vegetated habitats.

The focus of this study is the distribution of VAM plants and VAM propagules across the Pumice Plain of Mount St. Helens, and their relationship with microsites. The relationship between VAM and microsites is of interest because microsites are crucial to the colonization dynamics on Mount St. Helens (del Moral and Bliss 1993, Titus 1995).

During primary succession on volcanic substrates, it is unlikely that pioneer species would depend on mycorrhizae (Allen 1991). Only non-host and facultatively mycotrophic species could invade these sites. Obligately mycotrophic species would be prevented from establishing until a population of VAM fungi was present in the soil, presumably having arisen in association with facultatively mycotrophic species. Seral sequences may reflect the mycorrhizal dependence of the colonizing species (Allen 1991). Thus, the pattern of VAM distribution across the primary successional landscape of the Pumice Plain may regulate plant invasion patterns (Allen 1988). However, previous to this study, the distribution of VAM propagules across the Pumice Plain was unknown.

VAM propagules are composed of spores, hyphae and VAM colonized roots. The two indices of VAM density, spore counts and degree of root colonization, are not necessarily correlated (Louis and Lim 1987; Johnson et al. 1991). Spore counts assess only one type of propagule while colonization indirectly estimates all types of propagules. There-

fore, root colonization is a more accurate measure of total VAM density. In this study, the distribution of VAM has three facets: 1) The presence of VAM fungal propagules in the soil, i.e., the mycorrhizal inoculum potential (MIP), which is determined through a root bioassay; 2) the presence of VAM plants; and 3) the presence of VAM fungal spores in the soil. This study examines the distribution of these three components of VAM across the Pumice Plain.

## **METHODS**

Study site. The Pumice Plain was formed by the 18 May, 1980 eruption of Mount St. Helens (46°12'N, 122°11'W). The Pumice Plain covers 20 km² immediately north of the crater between 1150–1300 m elevation. It was formed by the deposit of up to 200 m of material from a debris avalanche, subsequent pyroclastic flows, air-fall pumice, and was repeatedly impacted by later lahars. The Pumice Plain is blanketed by pumice that ranges in depth from 10 to 200 m. The surface is flat or gently sloping, with numerous gullies created by erosion dissecting the surface. Surface pumice particles range in size from 1 mm to 10 cm (del Moral and Bliss 1993).

The climate is maritime, with cool wet winters and warm, dry summers. Periods of drought often occur in July and August. Annual precipitation averages 2373 mm, yet usually less than 5% of this falls between June and August. The growing season begins in June and ends by early September. Temperatures range from mean monthly minima of  $-4.2^{\circ}$ C in January and  $7.3^{\circ}$ C in August to maxima

<sup>&</sup>lt;sup>1</sup> Present address: Jonathan H. Titus, Department of Biological Sciences, University of Nevada, Las Vegas, Las Vegas, NV 89154-4004.

of 0.5°C in January and 22.2°C in July (Spirit Lake Ranger Station (987 m a.s.l.), Pacific Northwest River Basins Commission 1969). Summer temperatures range from 0 to 35°C with a mean ca. 12°C. Surface soil temperatures are often very high in the summer approaching 50°C on tephra surfaces (Reynolds and Bliss 1986).

Pumice Plain soil is immature, with very low concentrations of carbon, nitrogen, and microbial biomass (del Moral and Bliss 1993). Considerable variation in soil moisture values has been recorded between and within microsites. Substrates with fine particles contain more moisture than areas with coarse particles and erosion rills are more moist than other microsites (del Moral and Bliss 1993). In summer the surface tephra dries quickly between rains, thus, most Pumice Plain habitats do not remain moist for periods sufficient to allow seedling establishment. However, the surface layer of tephra acts as a mulch to impede evaporation and is capable of holding considerable moisture at lower depths so that adult plants rarely suffer from drought (Reynolds and Bliss 1986).

Microsites. The seven types of microsites in this study appear to differ in environmental characteristics on the spatial scale of an individual seed or seedling. They were chosen because personal observation and the literature both suggest them to be important to revegetation processes on the Pumice Plain. These sites are:

Flat—sites which have homogeneous gravel, sand or silt substrates in which the topography is level. Pumice particles are less than 5 cm in diameter. Flat sites occupy most of the Pumice Plain and are sparsely vegetated.

Rill—small gullies formed by erosive water action.

These are linear habitats that marginally protect seedlings from wind, collect more snow, and have lower solar radiation (del Moral and Bliss 1993). Rill edges are more stable than rill bottoms and drainages.

Near-rock—adjacent to rocks larger than 25 cm in diameter. On exposed surfaces rocks protect seedlings from direct solar exposure, reduce wind and surface temperatures, and are more likely to trap seeds.

Ridges—sites located on small ridgetops where there is evidence of extensive wind erosion.

Lupinus patch—sites associated with dense patches of living and dead Lupinus lepidus Douglas. These sites contain higher levels of soil nitrogen and lupines effectively trap seeds and organic matter. Lupinus patches are described in Halvorson et al. (1992), del Moral and Bliss (1993), del Moral et al. (1995), Bishop (1996).

Crowded vegetation—sites located in thick vegetation on new volcanically emplaced surfaces which are not dominated by *L. lepidus*.

Refugia—sites with pre-eruption soil exposed by erosion in which some belowground plant or-

gans survived and subsequently sprouted. Refugia are densely vegetated and are confined to the eastern fringe of the Pumice Plain on steep north facing slopes. Refugia vegetation is described by del Moral et al. (1995).

These sites were investigated to determine the distribution of VAM propagules and plants across the Pumice Plain. The first study looks at the distribution of VAM propagules, the second at the distribution of VAM plants and the third at the distribution of VAM spores.

Corn bioassay for assessing VAM propagule distribution. Soil samples were collected at 20 representative locations within each site (except refugia) in July 1991. Four 250 g samples from the upper 8 cm of soil were collected at each location and combined to form two composite samples. Soil was sifted to remove all particles larger than 4 mm, amended with 20% sterile perlite to increase porosity, and 650 g was placed into 10 cm by 10 cm freely draining plastic pots. Bioassays were conducted with non-fungicide treated Zea mays seeds. All pots were watered daily with tap water. Fertilizer was applied in 50 ml aliquots per pot of Colwell's solution minus phosphorus at planting and at weekly intervals throughout the experiment. Colwell's solution mimics natural proportions of nutrients in typical temperate soils (Colwell 1943; R. B. Walker personal communication). The control consisted of 20 pots of sterile greenhouse soil placed randomly among the treatment pots and planted with corn to determine if contamination by greenhouse VAM propagules occurred. Previous work showed that VAM propagules, if present, rapidly colonize corn in the greenhouse (Titus personal observation). Pots were randomized and maintained at the University of Washington Botany Greenhouse at 20-25°C, and rotated every 10 days. Bioassay plants were grown for 35 days from 20 July to 14 August 1991. Plants were harvested, roots washed, and frozen at -18°C until October 1991 at which time roots were assayed for VAM colonization. The quantity of inoculum in the soil, mycorrhizal inoculum potential (MIP), was estimated by percent colonization of corn roots (Moorman and Reeves 1979; Doerr et al. 1984; Johnson and McGraw 1988).

Staining. Roots were washed, cleared and stained with trypan blue (Brundrett et al. 1994; E. Cázares, Oregon State University, personal communication). Percent colonization was estimated by placing a grid of 1 cm squares below a petri plate which contained the root sample under a dissecting microscope. One hundred locations where a root crossed a line on the grid were scored for mycorrhizae. Many samples were examined under higher power to ascertain that the fungus was indeed VAM. Root segments containing vesicles, arbuscles or intercellular hyphal coils or hyphae were recorded as being colonized. The number of mycorrhizal "hits" is an

estimate of the percent of the root colonized (Brundrett et al. 1994).

Mycorrhizal colonization of pioneer species. The roots of 14 plants of six major pioneer species were collected from each of the seven site types at different locations across the Pumice Plain. Most root samples were collected during July, 1992, and the remainder during July, 1993. All sampled plants were at least 4 m from their nearest neighbor, except for those in lupine patches, crowded vegetation and refugia. Species sampled were Anaphalis margaritacea (L.) Benth. & Hook., Carex mertensii Prescott, Epilobium angustifolium L. ssp. circumvagum Mosq., Hieracium albiflorum Hook., Hypochaeris radicata L. and Penstemon cardwellii. In addition to these target species, roots were collected from several other naturally occurring species where they occurred. Roots of these species were not collected in all microsite types because they only occurred in certain ones. Roots were washed when harvested and stored in alcohol until they were cleared and stained to assess VAM colonization as above.

Spore isolation. Soil samples were collected using two sampling regimes. For the first regime soil was collected from 20 representatives of each of the seven site types. Four 100 ml soil samples were collected from the top 8 cm of soil and combined to form a single sample during August 1993.

The second sampling regime was part of a larger study. Percent cover of each plant species in 150 100 m<sup>2</sup> circular plots was assessed across the Pumice Plain during summer, 1993. Vegetation of these plots were grouped into five habitat types based on substrate and vegetation (del Moral et al. 1995). In conjunction with vegetation sampling, 100 ml of soil were collected at each of four locations within each plot and combined into a single sample.

Soil samples were dried at room temperature and stored at 3°C in sealed plastic bags. Spores were extracted from two subsamples of the soil from each plot by the wet-sieving and decanting technique (Gerdemann and Nicolson 1963; Pacioni 1992; Brundrett et al. 1994). One hundred and fifty ml of soil were placed into a 1.651 mm mesh sieve above 0.417 and 0.052 mm mesh sieves. The soil was washed vigorously with water. Roots in the top sieve and soil from the fine mesh bottom sieve were examined in a petri dish under a dissecting microscope at 40× power for VAM spores.

In order to compare spore extraction techniques, the soil from ten samples with two replicates each were analyzed using both the wet-sieving with decanting technique and the differential water/sucrose centrifugation method (Ianson and Allen 1986). Selected soil samples were those likely to contain VAM spores. Spore isolation efficiency was not improved using the differential water/sucrose technique. Although Ianson and Allen (1986) found better spore isolation with the differential water/

Table 1. VAM Corn Bioassay for the Determination of Mycorrhizal Inoculum Potential (MIP) of Pumice Plain Soil. Soil MIP is shown by percent VAM colonization of corn roots. % plants colonized is the percentage of the plants of each species which were colonized by VAM. (mean  $\pm$  standard deviation, n = 20 paired samples).

Microsite	MIP	% plants colonized
Flat	0	0
Near Rock	0	0
Ridge	0	0
Rill	$0.3 \pm 0.6$	15
Lupine Patch	$3.0 \pm 3.3$	60
Crowded Vegetation	$4.3 \pm 3.0$	70

sucrose technique, the extremely low organic matter content of Pumice Plain soils obviated the need for improved resolution in the case of these soils.

Spores were isolated and stored dry on filter paper. Spore types were determined from the experience of the third author and the use of spore identification guides (Mosse and Bowen 1968; Gerdemann and Trappe 1974; Trappe 1982; Morton 1988; Schenck and Perez 1990).

Data analysis. Mean percent mycorrhizal colonization was determined to yield MIP (Experiment I) and mean colonization (Experiment II). In Experiment III, spore density was averaged and richness determined for each site and for each habitat type. The preponderance of zeros precluded statistical data analysis, so values are reported only as observational data. Although both parametric and non-parametric techniques are robust for violations of their respective assumptions, the statistical techniques appropriate for analysis of this experimental design (e.g., one-way ANOVA or the Kruskal-Wallace test (Zar 1984)) are invalid for the analysis of data with many zeros. Even non-parametric statistics require homoscedastic variances. This aside, the patterns in the data are clearly apparent without statistical tests. Frequency of VAM colonization or spores are also reported.

#### RESULTS

Corn bioassay for assessing VAM propagule distribution. Ridge, flat, and near-rock substrates contained no detectable MIP. Rill microsites occasionally contained VAM propagules, whereas lupine patch and densely vegetated site soils usually contained mycorrhizal inoculum (Table 1).

Mycorrhizal colonization of pioneer species. Anaphalis margaritacea, Hieracium albiflorum and Hypochaeris radicata were without mycorrhizal colonization in flat, ridge and near-rock sites, but were mycorrhizal in rill, lupine patch, crowded vegetation and refugia (Table 2). Carex mertensii was without mycorrhizal colonization in all sites,

Table 2. VAM Colonization of Plant Species Collected from Microsites on the Pumice Plan. (mean  $\pm$  standard deviation, n = 14). Flat, ridge and near rock microsites contained no VAM plants and are not shown.

				Mi	crosite			
	Rill		Lupine pa	atch	Crowded veg	etation	Refugia	ì
Species	% VAM colonization	% plants colonized	% VAM colonization	% plants colo- nized	% VAM colonization	% plants colo-nized	% VAM colonization	% plants colo-nized
Anaphalis margaritacea	$1.1 \pm 2.6$	36	$8.3 \pm 15.3$	57	$10.2 \pm 12.6$	86	$6.4 \pm 7.8$	63
Carex mertensii	0	0	0	0	0	0	$0.1 \pm 0.4$	14
Epilobium angustifolium	0	0	$2.0 \pm 5.6$	29	$0.4 \pm 1.6$	14	$4.2 \pm 6.6$	36
Hieracium albiflorum	$0.2 \pm 0.8$	14	$4.9 \pm 9.3$	64	$5.3 \pm 12.3$	64	$8.9 \pm 10.5$	50
Hypochaeris radicata	$0.9 \pm 2.5$	29	$8.1 \pm 11.8$	64	$8.0 \pm 15.5$	33	$10.7 \pm 13.7$	50
Penstemon cardwellii	$0.8 \pm 2.8$	7	$3.2 \pm 4.2$	43	$7.0 \pm 7.8$	79	$15.7 \pm 9.0$	79

except for a trace of VAM in refugia. Epilobium angustifolium was not colonized in rill microsites, but was occasionally colonized in lupine patch, crowded vegetation and refugia. The species with the highest mycorrhizal colonization was A. margaritacea, and the site with the most VAM plants was crowded vegetation. Most of the VAM fungal hyphae observed were of the fine endophyte type. Non-target species were all nonmycorrhizal in flat, ridge and near-rock sites (Table 3). Only Juncus parryi Engelm. was mycorrhizal in rill microsites. Juncus parryi and Lupinus lepidus were mycorrhizal in lupine patches. Many species were mycorrhizal in crowded sites and refugia, and species restricted to refugia were usually mycorrhizal.

Spore distribution. No spores were found in flat, rill, or near-rock sites. Densities were variable where spores were found in dead lupine, crowded vegetation, and refugia (Table 4). Dead lupine and crowded vegetation microsites with VAM spores often were located far from refugia. Most refugia samples contained many spores.

Pumice barrens, pyroclastic surfaces and drainages (del Moral et al. 1995) rarely contained VAM fungal spores (Table 5). Samples containing spores were usually located near refugia. The only exception was an isolated barren pumice site which also contained a large willow. Lupine patches occasionally contained spores which, when present, were in high densities. Lupine patches which contained spores were widely spread across the Pumice Plain. Refugia almost always contained spores.

Three spore types were found: Glomus macro-carpum (Tul. and Tul.), Glomus mosseae (Nicol. and Gerd.) Gerd. and Trappe, and an Acaulospora type. The most common spores found were dead (which are empty), dark brown, brassy, or black. There was usually only one spore type present in a sample (not including dead spores which were usually present), however, all three spore types occurred in microsites and habitat types where spores were common.

## DISCUSSION

VAM distribution. Based on observational data only, microsites differ in MIP and spore density, and pioneer species differ in mycorrhizal colonization depending on the microsite it inhabits. Sites with thicker vegetation contained more VAM propagules and VAM plants.

After the 1980 eruption, the Pumice Plain was free of VAM fungal propagules (Allen 1987). The VAM propagules detected in this study show that dispersal forces, most likely animals (Allen 1987), have been returning VAM propagules to this landscape. The invasion of VAM propagules is sporadic as some microsites contain more mycorrhizal propagules and more heavily colonized plants than do other microsites. This supports the patch-dynamic model which proposes that the pattern of VAM fungal propagules dispersed by animals searching among patches for food and cover in sparsely vegetated landscapes creates a patchy distribution of inoculum (Allen 1987, 1988). This patchiness may also result from the ability of certain microsites to effectively trap windborne or waterborne propagu-

VAM spores were uncommon across the Pumice Plain, but they are present in crowded and lupine patch microsites. The bulk of the Pumice Plain appears to remain VAM spore free. Evidence for a non-patchy landscape level spread of VAM fungi was observed in the plots adjacent to refugia in which VAM propagules were found. Since refugia are on steep slopes, VAM propagules could immigrate to adjacent barren and drainage habitats by erosion. Plant diversity is also slightly higher in areas adjacent to refugia (del Moral et al. 1995). However, one pumice barren plot distant from sites with high levels of VAM spores contained VAM spores. This is evidence for a patchy distribution of VAM spores and adds support to the patch-dynamic model (Allen 1988). This pumice barren site contained a large willow which is probably a locus for small mammal activity in a barren landscape. In

Table 3. VAM Colonization of Plant Species Collected from Microsites on the Pumice Plain. n = sample size. (mean  $\pm$  standard deviation).

					Microsite	e			
		Flat			Ridge			Near rock	
Species	n	% VAM coloni- zation	% plants colo- nized	n	% VAM coloni- zation		n	% VAM coloniza-	% plants colo-nized
Achillea millefolium									
Agrostis pallens	4	0	O	2	0	0	4	0	0
Agrostis scabra	7	$2.3 \pm 6.1$	14	2	0	0	7	0	0
Blechnum spicant									
Calyptridium umbellatum	4	0	0				3	0	0
Carex pachystachya	3	0	0	3	0	0	3	0	0
Carex phaeocephala	3	0	0	3	0	0	3	0	0
Cirsium arvense							2	0	0
Epilobium anagallidifolium	5	0	0	2	0	0	4	0	0
Epilobium brachycarpum									
Epilobium ciliatum									
Eriogonum pyrolifolium									
Fragaria virginiana									
Gnaphalium uliginosum									
Juncus mertensianus									
Juncus parryi	10	0	0				2	0	0
Luetkea pectinata									
Lupinus latifolius	4	0	0				4	0	0
Lupinus lepidus	17	0	0	6	0	0	16	0	0
Luzula parviflora									
Phacelia hastata									
Polygonum minimum									
Ribes laxiflorum									
Rubus lasiococcus									
Rubus spectabilis									
Saxifraga ferruginea	5	0	0				4	0	0
Sambucus racemosa	-	_						_	_
Senecio sylvaticus	8	0	0	4	0	0	6	$0.7 \pm 1.6$	17
Smilicina racemosa	-	-	-		-		_		
Spergularia rubra	2	0	0	0	0	0	2	0	0
Vaccinium membranaceum	_	-		_	-	-	_	-	
Vancouveria hexandra									

addition, several of the crowded vegetation and lupine patch microsites which contained VAM spores were isolated across the Pumice Plain far from refugia, adding further support to the patch-dynamic model. The differences in spore counts for lupine patch and refugia between Tables 4 and 5 were not unexpected due to the large standard deviations and patchy nature of spore distribution (Anderson et al. 1983; St. John and Koske 1988).

# Mycotrophic Status of Colonizing Species

Glomus tenuis. Glomus tenuis (Greenall) Hall is distinguished by hyphal diameters in the 0.5–1.5 μm range (Hall 1987). Other Glomus species have coarse (5–30 μm in diameter) hyphae (McGonigle and Fitter 1990; Wang et al. 1993). Therefore, only root colonizations caused by G. tenuis can be identified confidently in the absence of sporulating structures (Carling and Brown 1982; Hall 1987). Colonization by G. tenuis has been found to be

highest in dry very low phosphorus environments (Rabatin 1979), low pH soils (Wang et al. 1993), and in alpine environments (Read and Haselwandter 1981; Mullen and Schmidt 1993). Glomus tenuis is also often the dominant VAM fungal species in pioneer species and disturbed environments (Daft and Nicolson 1974). In this study, fine endophyte hyphae were observed frequently, but no spores of G. tenuis were detected. Glomus tenuis spores may be too small (7-12 µm) to be extracted by the wet sieving technique (Hall 1987; Wang et al. 1993). Thus the possibility exists that these spores are common but were not detected. Although spores are the only way to identify Glomus species, they are not indicative of the actual infectivity of a soil and should be used only in conjunction with other indices. For example, no spores were detected in rill microsites, but there was some VAM colonization of corn roots and the target species in rills were occasionally VAM.

TABLE 3. EXTENDED

				Microsite				
	Rill	<u>.</u>		Lupine patch			Crowded vegetati	on
n	% VAM colonization	% plants colonized	n	% VAM colonization	% plants colonized	n	% VAM colonization	% plants colonnized
						2	6	100
4	0	0	4	0	0	12	$8.0 \pm 21.8$	25
7	0	0	7	0	0	6	$0.4 \pm 0.9$	17
2	0	0	1	0	0			
2 3 3	0	0	3	0	0	5	0	0
3	0	0	3	0	0	5	0	0
3	0	0	2 4	0	0	3	$5.7 \pm 4.0$	100
			4	0	0	2 2	0	0
2	0	0				2	$15.0 \pm 7.1$	100
						2	0	0
						1	0	0
						6	0	0
11	$0.5 \pm 1.5$	9	8	$9.0 \pm 7.0$	88	16	$0.8 \pm 2.6$	13
1	0	0						
						5	0	0
16	0	11	19	$0.2 \pm 0.6$	11	6	$0.7 \pm 1.6$	77
						14	0	0
						2 2	$4.0 \pm 5.7$	50
						2	0	0
5	0	0	1	0	0	2	$1.0 \pm 1.4$	50
4	0	0	5	0	0	4	$1.5 \pm 1.9$	50
2	0	0				0	0	0

Carex spp. are considered to be non-hosts (Powell 1975; Anderson et al. 1984), although mycorrhizal Carex spp. have been found in the alpine (Read and Haselwandter 1981; Allen et al. 1987) and in grasslands (Read et al. 1976). VAM Carex mertensii plants were only observed in two individuals in this study. Juncus parryi is generally thought to be in a non-host genus (Powell 1975). However, in this case it was heavily colonized by VAM in rill, lupine patch and crowded sites. Genera-wide generalizations of mycorrhizal dependence may be inaccurate and extensive examinations of the species must be conducted to ascertain mycotrophy (Read et al. 1976; Newman and Reddell 1987).

Annuals are often considered to be non-hosts (Trappe 1987; Boerner 1992; Peat and Fitter 1993), but in this survey *Senecio sylvaticus* was frequently mycorrhizal. Allen et al. (1992) found the annual *Epilobium paniculatum* to be mycorrhizal in a lupine patch. In this survey the species was found to be nonmycorrhizal.

Allen et al. (1992) found Lupinus latifolius J.

Agardh. and *L. lepidus* to be mycorrhizal, while this survey found *L. lepidus*, but not *L. latifolius*, to be mycorrhizal. Avio et al. (1990) observed *Lupinus* to be a strongly non-host genus. O'Dell and Trappe (1992) found both *L. lepidus* and *L. latifolius* to be occasionally mycorrhizal. They located a mycorrhizal *L. latifolius* on Mount St. Helens but did not find a mycorrhizal *L. lepidus* on the volcano. O'Dell and Trappe (1992) suggested that VAM fungi may need to be established on a companion host before colonizing roots of lupines.

Most plant species now colonizing Mount St. Helens barren sites appear to be facultatively mycotrophic (Titus 1995). This status supports a broad range of tolerance to VAM, from rarely mycorrhizal to nearly always colonized depending upon the species, neighboring species and site conditions (Allen 1991; Boerner 1992).

VAM fungal species. VAM fungal richness was low, with only three spore types, but greater than the single species (Glomus macrocarpum) found in the blast zone by Allen et al. (1984), Allen and

TABLE 3. CONTINUED

		Microsite	
		Refugia	
Species	n	% VAM colonization	% plants colo-nized
Achillea millefolium			
Agrostis pallens			
Agrostis scabra			
Blechnum spicant	2	$5.0 \pm 0$	100
Calyptridium umbellatum			
Carex pachystachya			
Carex phaeocephala			
Cirsium arvense			
Epilobium anagallidifolium			
Epilobium brachycarpum			
Epilobium ciliatum			
Ériogonum pyrolifolium			
Fragaria virginiana	3	$8.3 \pm 13.6$	67
Gnaphalium uliginosum			
Iuncus mertensianus			
Iuncus parryi	0	$0.5 \pm 1.5$	9
Luetkea pectinata	4	0	0
Lupinus latifolius			
Lupinus lepidus			
Luzula parviflora			
Phacelia hastata			
Polygonum minimum			
Ribes laxiflorum	4	$15.0 \pm 9.1$	100
Rubus lasiococcus	2	0	0
Rubus spectabilis	4	$16.0 \pm 5.9$	100
Saxifraga ferruginea			
Sambucus racemosa	4	$20.0 \pm 12.8$	100
Senecio sylvaticus	2	$4.0 \pm 2.8$	100
Smilicina racemosa	4	$30.0 \pm 14.1$	100
Spergularia rubra	•	23.0 = 11.1	100
Vaccinium membranaceum	4	$25.0 \pm 7.7$	100
Vancouveria hexandra	2	0	0

MacMahon (1988), and Allen et al. (1992). This indicates that VAM fungal species are invading the blast zone or at least proliferating into detectable densities. The preponderance of inviable spores found in this study is not unusual (Read et al. 1976;

Table 4. Number and Richness of VAM Fungal Spores in 150 ml Soil Samples from Microsites on the Pumice Plain. (mean  $\pm$  standard deviation for spore counts, n = 20 for each microsite type). \(^1\) Mean richness is based only on samples which contained spores.

		-	
Microsite	Mean number of spores	% samples with spores	Mean rich- ness <sup>1</sup>
Flat	0	0	
Near Rock	0	0	
Ridge	0	0	_
Rill	0	0	_
Lupine Patch	$13.6 \pm 29.2$	55	1.4
Crowded Vegetation	$18.4 \pm 41.1$	70	1.3
Refugia	$20.7 \pm 49.6$	85	1.8

Berliner and Torrey 1989). The patchy nature of VAM species distribution is evidenced by the large variance in spore densities and by the presence of different spore types in different sites with little overlap. However, each species was present in the microsites and habitat types which had detectable spore populations. It is important to note the difference in sampling intensity between above- and belowground environments. Plot size in del Moral et al. (1995) was 100 m<sup>2</sup>, where as the surface area of the belowground sampling effort was only approximately 400 cm<sup>2</sup>, which is 0.0004 as large as the aboveground sampling area. Therefore, statements about patchy spore distributions must be regarded in the light of the small belowground sampling area (Anderson et al. 1983). In the few studies which address VAM species distribution, richness is usually low and density variable. It is therefore difficult to draw conclusions about successional patterns in VAM fungal types from the results presented here.

# CONCLUSION

This study assessed both VAM colonization and VAM fungal propagules. The results are comple-

Table 5. Number and Richness of VAM Fungal Spores in 150 ML Soil Samples from Habitat Types on the Pumice Plain. Habitat types based on del Moral et al. (1995). "Near" indicates a site adjacent to a refugia, "far" indicates a site distant from a refugia. n = sample size. (mean  $\pm$  standard deviation for spore counts). <sup>1</sup> Mean richness is based only on samples which contained spores.

Habitat type	n	Mean number of spores	% samples with spores	Mean richness <sup>1</sup>
Pumice Barrens-near	11	$1.5 \pm 3.0$	4	1
Pumice Barrens-far	32	$0.03 \pm 0.2$	3	1
Pyroclastic Surfaces	15	0	0	_
Drainages-near	4	$0.3 \pm 0.5$	25	1
Drainages-far	15	0	0	_
Wetlands	23	0	0	_
Lupine Patches	16	$10.5 \pm 26.2$	25	1.5
Refugia	26	$14.5 \pm 33.1$	62	1.3

mentary and converge to the conclusion that the Pumice Plain remains essentially VAM free, except for the few isolated lupine patch and crowded sites. Refugia contain VAM fungal propagules and mycorrhizal plants. The sparse vegetation of the Pumice Plain is composed largely of facultatively mycotrophic species which are at present nonmycorrhizal.

#### ACKNOWLEDGMENTS

Thanks to Doug Ewing, Elaine Bassett, and Jeanette Milne for attending to the plants in the Greenhouse and for good advice. Extensive lab work was contributed by JoEllen VanDeMark. Thanks to Joseph Ammirati, Shannon Berch, Thomas Odell, and David Ianson for advice. Thanks to Priscilla Titus for editing the manuscript and for her generous encouragement and support. NSF Grants BSR-89-06544 and DEB-9406987 to Roger del Moral helped support this research.

# LITERATURE CITED

- ALLEN, E. B., J. C. CHAMBERS, K. F. CONNER, M. F. ALLEN, AND R. W. BROWN. 1987. Natural reestablishment of mycorrhizae in disturbed alpine ecosystems. Arctic and Alpine Research 19:11–20.
- ALLEN, M. F. 1987. Re-establishment of mycorrhizae on Mount St. Helens: migration vectors. Transactions of the British Mycological Society 8:413–417.
  - . 1988. Re-establishment of VA mycorrhizae following severe disturbance: comparative patch dynamics of a shrub desert and subalpine volcano. Proceedings of the Royal Society of Edinburgh 94B:63-71.
  - ——. 1991. The Ecology of Mycorrhizae. Cambridge University Press, New York.
- —, C. CRISAFULLI, C. F. FRIESE, AND S. L. JEAKINS. 1992. Re-formation of mycorrhizal symbioses on Mount St. Helens, 1980–1990: interactions of rodents and mycorrhizal fungi. Mycological Research 96: 447–453.
- AND J. A. MACMAHON. 1988. Direct VA mycorrhizal inoculation of colonizing plants by pocket gophers (*Thomomys talpoides*) on Mount St. Helens. Mycologia 80:754–756.
  - —, J. A. MACMAHON, AND D. C. ANDERSEN. 1984. Reestablishment of Endogonaceae on Mount St. Helens: survival of residuals. Mycologia 76:1031–1038.
- Anderson, R. C., A. E. Liberta, and L. A. Dickman. 1984. Interactions of vascular plants and vesicular-arbuscular mycorrhizal fungi across a soil moisture-nutrient gradient. Oecologia 64:111–117.
- ——, A. E. LIBERTA, L. A. DICKMAN, AND J. A. KATZ. 1983. Spatial variation in vesicular-arbuscular mycorrhiza spore density. Bulletin of the Torrey Botanical Club 110:519–525.
- AVIO, L., C. SBRANA AND M. GIOVANNETTI. 1990. The response of different species of *Lupinus* to VAM endophytes. Symbiosis 9:321–323.
- Berliner, R. and J. G. Torrey. 1989. Studies on mycorrhizal associations in Harvard Forest, Massachusetts. Canadian Journal of Botany 67:2245–2251.
- BISHOP, J. G. 1996. Demographic and population genetic variation during colonization by the herb *Lupinus lepidus* on Mount St. Helens. Ph.D. dissertation. University of Washington, Seattle, WA.
- BOERNER, R. E. J. 1992. Plant life span and response to

- inoculation with vesicular-arbuscular mycorrhizae. I. Annual versus perennial grasses. Mycorrhiza 1:153–161
- Brundrett, M. C., L. Melville, and L. Peterson. 1994. Practical methods in mycorrhiza research. Mycologue Publications, Guelph, Ontario.
- Carling, D. E. and M. F. Brown. 1982. Anatomy and physiology of vesicular-arbuscular and nonmycorrhizal roots. Phytopatology 72:1108–1114.
- Colwell, W. E. 1943. A biological method for determining the relative boron content of soils. Soil Sciences 56:71–94.
- DAFT, M. J. AND T. H. NICOLSON. 1974. Arbuscular mycorrhizas in plants colonizing coal wastes in Scotland. New Phytologist 73:1129–1138.
- DEL MORAL, R. AND L. C. BLISS. 1993. Mechanisms of primary succession: insights resulting from the eruption of Mount St. Helens. Advances in Ecological Research 24:1–66.
- DEL MORAL, J. H. TITUS, AND A. M. COOK. 1995. Early primary succession on Mount St. Helens, Washington, USA. Journal of Vegetation Science 6:107–120.
- DOERR, T. B., E. F. REDENTE, AND F. B. REEVES. 1984. Effects of soil disturbance on plant succession and levels of mycorrhizal fungi in a sagebrush-grassland community. Journal of Range Management 37:135–139.
- GERDEMANN, J. W. AND T. H. NICOLSON. 1963. Spores of mycorrhizal Endogone species extracted from soil by wet sieving and decanting. Transactions of the British Mycological Society 46:235–244.
- GERDEMANN, J. W. AND J. M. TRAPPE. 1974. Endogonaceae of the Pacific Northwest. Mycologia Memoir 5:1–76.
- HALL, I. R. 1987. Taxonomy and identification of vesicular-arbuscular mycorrhizal fungi. Angewandte Botanik 61:145–152.
- HALVORSON, J. J., E. H. FRANZ, J. L. SMITH, AND R. A. BLACK. 1992. Nitrogenase activity, nitrogen fixation, and nitrogen inputs by lupines at Mount St. Helens. Ecology 73:87–98.
- IANSON, D. C. AND M. F. ALLEN. 1986. The effects of soil texture on extraction of vesicular-arbuscular mycorrhizal fungal spores from arid sites. Mycologia 78: 164–168.
- JOHNSON, N. C. AND A. McGRAW. 1988. Vesicular-arbuscular mycorrhizae in taconite tailings. I. Incidence and spread of Endogonaceous fungi following reclamation. Agriculture, Ecosystems and Environment 21:135–142.
- ——, F. L. PFLEGER, R. K. CROOKSTON, S. R. SIMMONS, AND P. J. COPELAND. 1991. Vesicular-arbuscular mycorrhizae respond to corn and soybean cropping history. New Phytologist 117:657–663.
- Louis, I. and G. Lim. 1987. Spore density and root colonization of vesicular-arbuscular mycorrhizas in tropical soil. Transactions of British Mycological Society 88:207–212.
- MCGONIGLE, T. P. AND A. H. FITTER. 1990. Ecological specificity of vesicular-arbuscular mycorrhizal associations. Mycological Research 94:120–122.
- MOORMAN, T. AND F. B. REEVES. 1979. The role of endomycorrhizae in revegetation practices in the semi-arid west. II. A bioassay to determine the effect of land disturbance on endomycorrhizal populations. American Journal of Botany 66:14–18.
- Morton, J. B. 1988. Taxonomy of VA mycorrhizal fungi: classification, nomenclature, and identification. Mycotaxon 32:267–324.

- Mosse, B. and G. D. Bowen. 1968. A key to the recognition of some Endogone spore types. Transactions of the British Mycological Society 51:469–483.
- MULLEN, R. B. AND S. K. SCHMIDT. 1993. Mycorrhizal infection, phosphorus uptake, and phenology in *Ranunculus adoneus*: implications for the functioning of mycorrhizae in alpine systems. Oecologia 94:229–234.
- NEWMAN, E. I. AND P. REDDELL. 1987. The distribution of mycorrhizas among families of vascular plants. New Phytologists 106:745–751.
- O'DELL, T. E. AND J. M. TRAPPE. 1992. Root endophytes of lupin and some other legumes in Northwestern USA. New Phytologist 122:479–485.
- Pacific Northwest River Basins Commission. 1969. Climatological Handbook Columbia Basin States, Vol. 1, Part A and Vol. 2. Meterology Committee, Pacific Northwest River Basins Commission, Portland, OR.
- Pacioni, G. 1992. Wet-sieving and decanting techniques for the extraction of spores of vesicular-arbuscular fungi. Pp. 317–322 *in* J. R. Norris, D. J. Read, and A. K. Varma (eds.), Methods in Microbiology, Vol. 24. Academic Press, London.
- PEAT, H. J. AND A. H. FITTER. 1993. The distribution of arbuscular mycorrhizas in the British flora. New Phytologist 845–854.
- POWELL, C. L. 1975. Rushes and Sedges are non-mycotrophic. Plant and Soil 42:481–484.
- RABATIN, S. C. 1979. Seasonal and edaphic variation in vesicular-arbuscular mycorrhizal infection of grasses by Glomus tenuis. New Phytologist 83:95–102.
- READ, D. J. AND K. HASELWANDTER. 1981. Observations on the mycorrhizal status of some alpine plant communities. New Phytologist 88:341–352.

- ———, H. K. KOUCHEKI, AND J. HODGSON. 1976. Vesicular-arbuscular mycorrhiza in natural vegetation systems. I. The occurrence of infection. New Phytologist 76:641–653.
- REYNOLDS, G. D. AND L. C. BLISS. 1986. Microenvironmental investigations of tephra covered surfaces at Mount St. Helens. Pp. 147–152 *in* S. A. C. Keller (ed.), Mount St. Helens: five years later. Eastern Washington State University Press, Cheney, WA.
- St. John, T. V. and R. E. Koske. 1988. Statistical treatment of Endogonaceous spore counts. Transactions of the British Mycological Society 91:117–121.
- SCHENCK, N. C. AND Y. PEREZ. 1990. Manual for the identification of VA mycorrhizal fungi, 3rd ed. Synergistic Publications, Gainesville, FL.
- Titus, J. H. 1995. The role of mycorrhizae and microsites in primary succession on Mount St. Helens. Ph.D. dissertation. Department of Botany, University of Washington, Seattle, WA.
- Trappe, J. M. 1982. Synoptic Keys to the genera and species of zygomycetous mycorrhizal fungi. Phytopathology 72:1100–1108.
- . 1987. Phylogenetic and ecologic aspects of mycotrophy in the angiosperms from an evolutionary standpoint. Pp. 2–25 in G. R. Safir (ed.), Ecophysiology of VA mycorrhizal plants. CRC Press, Boca Raton, FL.
- WANG, G. M., D. P. STRIBLEY, P. B. TINKER, AND C. WALK-ER. 1993. Effects of pH on arbuscular mycorrhiza I. Field observations on the long-term liming experiments at Rothamsted and Woburn. New Phytologist 124:465–472.
- ZAR, J. H. 1984. Biostatistical analysis, 2nd ed. Prentice-Hall, Inc., Englewood Cliffs, New Jersey.